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### **ORIGINAL ARTICLE**

# Polyethylene glycol functionalized Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) for the efficient removal of heavy metals from *Ligusticum chuanxiong* Hort



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#### KEYWORDS

*Ligusticum chuanxiong* Hort; Heavy metals; Removal; Magnetic nanomaterial; MOFs; Polyethylene glycol

Abstract In recent years, the quality and safety issues of Chinese medicinal herbs have received great attention worldwide. Thereinto, heavy metal contamination has been one of the most serious concerns. Compared to the wide research in the analysis of heavy metals in medicinal herbs, the studies on the removal of heavy metals are relatively limited. In this study, polyethylene glycol functionalized Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) (Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG) was designed and synthesized to remove heavy metals from the decoction of Ligusticum chuanxiong Hort. The in-house fabricated Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG was characterized by a porous structure and a large specific surface area. Then, the efficiency of the material for the removal of five heavy metals was tested under optimal adsorption conditions. Meanwhile, the content of Senkyunolide A, Senkyunolide I, and Ferulic acid, the solid content, and the HPLC fingerprints similarity were used as the quality monitoring indicators of Ligusticum chuanxiong Hort decoction before and after the heavy metal removal. Results showed that the magnetic nanomaterial had excellent removal efficiency for As<sup>5+</sup> (81.4 %),  $Cd^{2+}$  (88.19 %), and  $Pb^{2+}$  (83.79 %) and certain removal efficiency for Ni<sup>2+</sup> (51.59 %) and  $Zn^{2+}$  (55.4 %) under the spiked concentration of 50 µg/mL. The content of Senkyunolide A, Senkyunolide I, and Ferulic acid were decreased by less than 8.00 % after the removal of heavy metals. Besides, the loss rate of solid content was only 0.18 %, and the finger-

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1878-5352 © 2023 The Author(s). Published by Elsevier B.V. on behalf of King Saud University. This is an open access article under the CC BY-NC-ND license (http://creativecommons.org/licenses/by-nc-nd/4.0/). prints similarity was over 99.9 %. The results indicated that  $Fe_3O_4$ @MIL-101(Cr)@PEG could efficiently and selectively remove heavy metals from *Ligusticum chuanxiong* Hort without affecting its effective components. Due to the advantages of low-cost, simple manipulation, and good efficiency, the material can be recommended for heavy metals removal from the aqueous solutions of medicinal herbs, providing a new and promising application for the removal of exogenous contaminants in medicinal herbs.

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#### 1. Introduction

Heavy metals have toxic effects on the body's metabolism and normal physiological functions, some of which have been recognized as harmful trace elements to the human body, such as Pb, Cd, Hg, and As. Excessive consumption of heavy metals in the human body can lead to various diseases, such as renal failure, chronic toxic symptoms, liver injury, and even cancer(Azimi et al., 2017). As common extrinsic contaminants in medicinal herbs, heavy metals not only affect the quality and safety of medicinal herbs but also hinder the long-term development of the related industry(Yanhong et al., 2017). At present, the content of heavy metals has become one of the important indicators of controlling the quality and safety of medicinal herbs(Coghlan et al., 2015; Lin et al., 2018). The contamination during planting and processing causes the accumulation of heavy metals in some medicinal herbs, which even exceed the maximum residue limits (MRLs) (Xionghai & Yitong, 2004). Some research shows that the excessive content of heavy metals in Chinese medicinal materials is widespread (Liang & Liang, 2021). For example, the total over-limit ratio of heavy metals was 22 % in 100 commonly used traditional Chinese medicines (Feng et al., 2014). Thereinto, the over-limit ratio of Cd, Pb, As, and Hg was 19 %, 5 %, 2 %, and 1 %, respectively. According to the statistical analysis of the literature about heavy metals pollution in traditional Chinese medicine from 2000 to 2016, the over-limit ratio of Pb, As, Cd, and Hg in traditional Chinese medicine was 3.46 %, 4.03 %, 2.91 %, and 1.41 %, respectively(Guo et al., 2017). Therefore, the pollution of heavy metals in traditional Chinese medicine was still one of the serious problems regarding its quality and safety. Most of the studies presently focused on the detection of heavy metals in medicinal herbs(Zhang et al., 2021), while limited ones were about the removal of heavy metals. It is urgent to establish effective methods for the removal of heavy metals in medicinal herbs to improve the quality of traditional Chinese medicine and ensure its sustainable development.

Up to now, numerous techniques have been used to remove heavy metals from water and other matrices, such as chemical precipitation, adsorption, biosorption, ion exchange, coagulation-flocculation, membrane separation, and supercritical fluid extraction(Fu & Wang, 2011; Turhanen et al., 2015; Yang et al., 2019). Thereinto, adsorption is one of the most commonly used methods. In this way, different adsorbents have been developed and applied, including carbon-based nanomaterials, molecular imprinting polymers, chitosan, magnetic nanoparticles, metal-organic frameworks (MOFs), etc. MOFs, a new kind of crystalline porous material formed by metal ions/clusters and multidentate organic ligands, are emerging as an attractive class of adsorbents(Zhou et al., 2012). They are generally characterized by open frameworks with accessible channels and cage structures(Jalayeri et al., 2020; Xiaolei Zhang et al., 2020). Due to their extraordinary properties, represented by high surface area, porous structures, tunable pores, availability of in-pore and outer-surface functionalization, and excellent mechanical stability, MOFs show great potential for the adsorption of hazardous pollutants(Hasan et al., 2016; Shi et al., 2018; Wang et al., 2015). MIL-101(Cr) is a self-assembly of Cr<sup>3+</sup> ions and an organic ligand (terephthalic acid). Compared with other MOFs, MIL-101(Cr) has a mesoporous molecular sieve structure and a large number of unsaturated Cr(III) sites, which could effectively bind with

electron-rich functional groups(Yang & Yan, 2011). In addition, MIL-101(Cr) is relatively stable under acidic and neutral conditions, which overcomes the shortcomings of water instability of some other MOFs such as MOF-5 and MOF-177, thus widening their applications in the pretreatment of traditional Chinese medicine(Jia et al., 2020). Although novel materials have been extensively investigated for environmental samples, limited studies have been conducted to remove heavy metals in medicinal herbs(Guo et al., 2020; Yanhong et al., 2017; Yufang & Danying, 2005; Xuejie Zhang et al., 2020). One of the main reasons might be the complex existing forms of heavy metals in medicinal herbs, such as complexation or embedding with matrix components. Besides, other problems can easily occur in the removal process, including secondary pollution, low adsorption capacity/efficiency, the difficulty of large-scale production, strict application requirements, and influences on effective components of medicinal herbs. Therefore, more studies are still needed to investigate adsorbents that can achieve good removal efficiency of heavy metals from medicinal herbs and simultaneously maintain their active ingredients.

This study aims to synthesize a magnetic nanomaterial to remove heavy metals from the decoction of Ligusticum chuanxiong Hort (L. chuanxiong Hort). The value of L. chuanxiong Hort ranked seventh in the export of Chinese herbal medicines, accounting for 2.68 % of the total export(Delin et al., 2019). As one of the representative medicinal herbs, L. chuanxiong Hort plays an essential role in the pharmaceutical industry with its abundant chemical components and unique pharmacological effects. It can be used to treat cardiovascular, cerebrovascular, respiratory, urinary, immune system, and gynecological diseases. Although L. chuanxiong Hort has unique properties in disease treatment and health care, the worldwide popularity of L. chuanxiong Hort has been hindered by the contamination of heavy metals mainly from polluted water, air, and soil(Fuyou et al., 2003; InSil et al., 2017). More than 30 literatures indicated the particularly severe contamination of cadmium in L. chuanxiong Hort(Delin et al., 2019). Since nanostructured composites have shown a good application prospect in heavy metals removal from plant extracts(Yongdi et al., 2019), the combination of magnetic nanoparticles and MOFs has been considered to realize the efficient removal of heavy metals from the decoction of L. chuanxiong Hort. Magnetic composites based on MOFs can inherit the excellent properties of MOFs and the rapid separation and recovery of magnetic material, thus avoiding the problems of secondary pollution or difficult separation/recycling of the adsorbents. As one of the most prominent MOFs, MIL-101(Cr), a chromium terephthalate solid characterized by cubic structure(Hupp & Poeppelmeier, 2005), has shown excellent stability in the air, water, and a variety of organic solvents(Burtch et al., 2014). Hence, this study aims to synthesize an adsorbent based on Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) and evaluate its performance for heavy metals (Pb, Cd, As, Ni, and Zn ions) removal.

Considering the physicochemical property of these components, functionalization of  $Fe_3O_4$ @MIL-101(Cr) by polyethylene glycol (PEG) was intended to block the adsorption between the adsorbents and the effective components of *L. chuanxiong* Hort. PEG is a nanomaterial that is widely used to construct biosafety nanocarriers and functionalize other materials(Xuejie Zhang et al., 2020), which can improve the hydrophilicity and the dispersibility of the materials. Therefore, the introduction of PEG in the adsorbents was expected to guarantee sufficient contact between heavy metals and trapping

media and play as a barrier to separating the main effective components of *L. chuanxiong* Hort from the magnetic nanomaterial (Ostolska & Wisniewska, 2017). To the best of our knowledge, the PEG functionalized Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) (Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) @PEG) was applied to remove heavy metals from medicinal herbs for the first time. Specifically, the physicochemical properties of Fe<sub>3</sub>-O<sub>4</sub>@MIL-101(Cr)@PEG were investigated. After the optimization of adsorption conditions, the adsorption mechanism was investigated by adsorption isotherm, kinetic and thermodynamic of heavy metals on Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG. Finally, the feasibility and reliability of heavy metals removal from the decoction of *L. chuanxiong* Hort were further studied and evaluated.

#### 2. Materials and methods

#### 2.1. Chemicals and materials

Iron chloride hexahvdrate (FeCl<sub>3</sub>·6H<sub>2</sub>O, 99.0 %), chromium (III) nitrate nonahydrate (Cr(NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O, 99.0 %), dopamine hydrochloride (C<sub>8</sub>H<sub>11</sub>NO<sub>2</sub>·HCl, 98.0 %), tris(hydroxymethyl) aminomethane ( $C_4H_{11}NO_3$ ,  $\geq 99.8$  %), sodium hydroxide (NaOH, 97.0 %), terephthalic acid (C8H6O4, 99.0 %), and phosphoric acid (H<sub>3</sub>PO<sub>4</sub>, ≥99.0 %) were purchased from Aladdin Biochemical Technology Co., ltd. (Shanghai, China). Sodium acetate ( $C_2H_3NaO_2$ , 99.0 %), ethylene glycol (C<sub>2</sub>H<sub>6</sub>O<sub>2</sub>, 98.0 %), and N, N-dimethylformamide (DMF, 99.5 %) were from Macklin Biochemical Co., ltd. (Shanghai, China). Polyethylene glycol (PEG, 600 g/mol) was purchased from J&K Scientific Technology Co., ltd. (Beijing, China). Hydrochloric acid (HCl,  $36.0 \sim 38.0$  %) and ethanol (C<sub>2</sub>H<sub>5</sub>OH, 99.7 %) were supplied by Chron Chemicals Co., ltd. (Chengdu, China). Nitric acid (HNO<sub>3</sub>, 70 %) was purchased from Crystal Clear Chemical Co., ltd. (Suzhou, China). Acetonitrile (ACN, HPLC grade) was purchased from Sigma-Aldrich (Shanghai, China). Distilled water was purified by a Milli-Q system (Millipore Co.).

Arsenic standard stock solution (As<sup>5+</sup>, 1000 µg/mL, 2 % HNO<sub>3</sub>), nickel standard stock solution (Ni<sup>2+</sup>, 1000 µg/mL, 5 % HNO<sub>3</sub>), zinc standard stock solution ( $Zn^{2+}$ , 1000 µg/ mL, H<sub>2</sub>O), cadmium standard stock solution (Cd<sup>2-</sup> 1000 µg/mL, 5 % HNO<sub>3</sub>), lead standard stock solution  $(Pb^{2+}, 1000 \ \mu g/mL, 5 \ \% \ HNO_3)$  were purchased from Henan R&D Center for certified reference material. Senkyunolide I (C12H16O4, Sichuan Naturewill biotechnology Co., ltd, 2203185901), Senkyunolide A (C<sub>12</sub>H<sub>16</sub>O<sub>2</sub>, Purifa technology development Chengdu Co., ltd, PRF8031301), and Ferulic acid (C<sub>10</sub>H<sub>10</sub>O<sub>4</sub>, China Institute for identification of pharmaceutical and biological products, 110773–200611) were applied with the purity above 99.0 %. L. chuanxiong Hort samples were purchased from local pharmacies in Chengdu, China.

#### 2.2. Instrument conditions

The analysis of heavy metals was performed by an inductively coupled plasma-optical emission spectrometer (ICP-OES, Agilent 5110). The following wavelengths of the measured elements were used: 188.980 nm (As), 231.604 nm (Ni), 213.857 nm (Zn), 214.439 nm (Cd), 220.353 nm (Pb). Other parameters of the ICP-OES were set as follows: forward power, 1.2 kW; plasma argon flow rate, 12.00 L/min; auxiliary

argon flow rate, 1.0 L/min; nebulizer argon flow rate, 0.7 L/min; read time, 10 s; stabilization time; 15 s.

The determination of active ingredients in the decoction of *L. chuanxiong* Hort was performed on a high-performance liquid chromatography system (HPLC, Thermo UltiMate 3000) connected with an ultraviolet detector. The chromatographic separation was realized by a Thermo BDS C18 column (4.6 mm  $\times$  250 mm, 5 µm). The mobile phase consisted of (A) acetonitrile and (B) water (containing 0.5 % phosphoric acid). Analyses were carried out with a gradient elution as follows: 10 %-16 % A at 0–15 min, 16 %-30 % A at 15–30 min, 30 %-40 % A at 30–35 min, 40 %-55 % A at 35–70 min. The detection wavelength was 276 nm. The injection volume was 10 µL. The column temperature and flow rate were set as 30 °C and 1.0 mL/min, respectively.

The in-house fabricated adsorbent was characterized by scanning electron microscopy (SEM, Hitachi SU8020), transmission electron microscopy (TEM, FEI Tecnai G2 F30), Fourier transform infrared spectroscopy (FT-IR, Nicolet IS10), X-ray diffraction (XRD, BRUCKER D8 ADVANCE) patterns, Vibrating Sample Magnetometer (VSM, BKT-4500), and specific surface area and porosity analyzer (Micromeritics ASAP 2460).

#### 2.3. Synthesis of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG

Polydopamine (PDA) has attracted much attention as a functional monomer owing to its unique coating ability and versatility (Duan et al., 2018). It can be selfpolymerized onto magnetic nanoparticles under mild conditions to prepare adherent PDA films (Zhang et al., 2012; Zhao et al., 2015). The film with uniform thickness can be formed on various materials through the strong interaction with catechol and amine groups of dopamine(Bai et al., 2017). Therefore, PDA-coated magnetic nanoparticles were prepared as the substrate of MOFs in this study. The synthesis of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG was referenced to our last work with some modification(Tang et al., 2022). Specifically, the synthetic route of magnetic MOF composites was as follows (Fig. 1): firstly, Fe<sub>3</sub>O<sub>4</sub> magnetic microspheres were prepared by the solvothermal method according to the reference(Shao et al., 2012). Then, Fe<sub>3</sub>O<sub>4</sub> microspheres were coated by a PDA layer via the polymerization of dopamine in the Tris-HCl buffer solution (pH = 8.5). The prepared Fe<sub>3</sub>O<sub>4</sub>@PDA was dissolved in deionized water containing a certain amount of Cr(NO<sub>3</sub>)<sub>3</sub>-.9H<sub>2</sub>O, terephthalic acid, and sodium acetate. The mixture was subsequently sealed in a Teflon-lined bomb and heated under 210 °C for 10 h. In this one-pot synthesis process, Cr(NO<sub>3</sub>)<sub>3</sub>·9H<sub>2</sub>O provided metal nodes while terephthalic acid served as organic building blocks. The obtained Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) were successively washed with DMF, soaked in hot ethanol for 5 h, washed several times with ethanol, and dried under vacuum at 150 °C. After that, these dehydrated nanoparticles were suspended in PEG solution (1.5 mg/mL in water) and stirred at 30 °C for 4 h. The functionalized nanoparticles were recovered by applying an external magnet and washed twice with water to remove un-grafted PEG. Finally, the adsorbents were washed with ethanol and dried under vacuum at 30 °C.



Fig. 1 Schematic of the preparation of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG and its application for heavy metals removal.

#### 2.4. Preparation of the decoction of L. chuanxiong Hort

*L. chuanxiong* Hort samples were randomly collected from local pharmacies. All samples were stored in sealed plastic bags below 4 °C for further use. The decoction of *L. chuanxiong* Hort was prepared as follows: 2.0 g of *L. chuanxiong* Hort was weighed accurately into a 250 mL polytetrafluoroethylene crucible. Then, 100 mL of ultrapure water was added to the crucible, boiled for 1 h, and separated from the samples. The last procedure was repeated twice. Finally, the combined decoction was filtered and concentrated under reduced pressure to 0.02 g/mL.

#### 2.5. Adsorption behavior study

Adsorption kinetics, isotherms, and thermodynamic experiments were performed to evaluate the adsorption behavior of  $Fe_3O_4@MIL-101(Cr)@PEG$  toward  $As^{5+}$ ,  $Ni^{2+}$ ,  $Zn^{2+}$ ,  $Pb^{2+}$ , and  $Cd^{2+}$  (Hayati et al., 2012; Xiong, Xue, et al., 2023). Specifically, the adsorption kinetics were carried out with 40 mg adsorbents in 50 mL of metal ion solutions (50 µg/mL of Pb, Cd, As, Ni, and Zn ions, pH = 5.36). After being shaken for a specific time interval (30–360 min), the kinetic factors were computed. The experimental adsorption data were modeled with the pseudo-first-order and pseudo-second-order kinetic models (Eq. (1) and (2), respectively):

$$\log(Q_e - Q_t) = \log Q_e - \frac{K_1}{2.303}t$$
 (1)

$$\frac{t}{Q_t} = \frac{t}{Q_e} + \frac{1}{K_2 Q_e^2} \tag{2}$$

where  $Q_e$  (mg/g) and  $Q_t$  (mg/g) are the amounts of the adsorbed heavy metals on Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG at equilibrium and different times (min), respectively; K<sub>1</sub> (h<sup>-1</sup>) and K<sub>2</sub>(g·mg<sup>-1</sup>·min<sup>-1</sup>) are the adsorption rate constants for the pseudo-first-order adsorption and pseudo-second-order adsorption, respectively.

The adsorption isotherms were carried out by adding 40 mg adsorbents into 50 mL of metal ions solution with various concentrations (0.8–100  $\mu$ g/mL, PH = 5.36). Two well-known Langmuir and Freundlich isotherm models were applied to analyze the isotherm data (Eq. (3) and Eq. (4), respectively):

$$\frac{C_e}{Q_e} = \frac{C_e}{Q_m} + \frac{1}{K_L Q_m} \tag{3}$$

$$\log Q_e = \frac{1}{n_F} \log C_e + \log K_F \tag{4}$$

where  $Q_e$  is the quantity of the adsorbed heavy metals per gram of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG at equilibrium (mg/g);  $Q_m$  is the maximum specific amount of heavy metals per gram of Fe<sub>3</sub>-O<sub>4</sub>@MIL-101(Cr)@PEG (mg/g); C<sub>e</sub> is the residual heavy metals concentration at equilibrium (mg/L);  $K_L$  is the Langmuir equilibrium constant (L/mg);  $K_F$  and n are indicators of adsorption capacity and adsorption intensity, respectively.

The thermodynamic adsorption parameters were also measured at different temperature (20 °C, 30 °C, and 40 °C). The mathematical relationship between  $\Delta G$ ,  $\Delta H$  and  $\Delta S$  were shown in Eq. (5). The thermodynamic linear relationship was shown in Eq. (6). The values of  $\Delta H$  and  $\Delta S$  can be calculated according to the intercept and slope of the fitting curve.

$$\Delta G = \Delta H - T \Delta S \tag{5}$$

 $lnK = \frac{\Delta S}{R} - \frac{\Delta H}{RT} \tag{6}$ 

where  $\Delta G (kJ \cdot mol^{-1})$ ,  $\Delta H (kJ \cdot mol^{-1})$ , and  $\Delta S (J \cdot mol^{-1} \cdot K^{-1})$  represent Gibbs free energy, enthalpy change, and entropy change of heavy metal ions adsorbed by the adsorbent, respectively; K is the thermodynamic equilibrium constant; T (K) is the temperature in Kelvin.

## 2.6. Removal and analysis of heavy metals in the decoction of L. chuanxiong Hort

As shown in Fig. 1, 40 mg of adsorbents were added to 50 mL of *L. chuanxiong* Hort decoction. The introduction of magnetism in the material allowed us to separate the adsorbent from the decoction of *L. chuanxiong* Hort by an external magnetic field. Therefore, the adsorbents were withdrawn by an external magnetic field after being shaken for a specific time interval. Then, the decoction was entirely transferred into a polytetrafluoroethylene crucible and digested with 0.5 mL nitric acid for 2 h. After digestion, the solution was further heated and concentrated to about 1 mL. Finally, the obtained concentrate was entirely transferred to a 10 mL volumetric flask and diluted with 2 % nitric acid solution to volume, which was ready for ICP-OES analysis. To compare the adsorption effect of heavy metal ions, the same experiments were conducted with Fe<sub>3</sub>O<sub>4</sub> and MIL-101(Cr), respectively.

#### 2.7. Reusability study

The regeneration and reusability of an adsorbent are imperative for its economic and practical applications. Thus, the adsorption-desorption cycles were repeated for some times to examine the regeneration efficiency of Fe<sub>3</sub>O<sub>4</sub>@MIL-101 (Cr)@PEG. With the method described in section 2.6, the adsorption process was performed by adding 40 mg of adsorbents into 50 mL of spiked decoction with a concentration of 50 µg/mL and an adjusted pH value. After adsorption, the desorption of Pb, Cd, As, Ni, and Zn ions from the adsorbents was carried out with the aqueous HCl solution (0.1 M) as eluent. The Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG nanoparticles were stirred in the above solution for 1 h at room temperature. Then, the adsorbents were washed with distilled water several times for the next adsorption-desorption cycle. At the same time, the material after the first desorption was dried at 60 °C and weighed to obtain the recovery rate of the adsorbent.

#### 2.8. Stability study

The stability of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG nanoparticles against acidic solutions is vital for long-time practical

applications. Therefore, in the present work, the structural stability of this adsorbent against HCl (pH = 1) was studied for 10 h. Briefly, 100 mg Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG nanoparticles were added into 50 mL acidic, and the mixture was stirred at room temperature for 10 h (Ahmadijokani et al., 2021). After that, the nanoparticles were collected, washed with chloroform, and dried at 60 °C. Finally, the XRD diffractograms of the pristine adsorbents, the adsorbents after heavy metal adsorption, and the acid-treated adsorbents were compared to evaluate their structural stability.

#### 2.9. Metal leaching rate

The adsorption process was performed by adding 40 mg of adsorbents into 50 mL of aqueous solution. It should be noted that no heavy metal standard solution was added in this process. After the same operation described in section 2.6, Fe and Cr ions were analyzed by ICP-OES.

#### 2.10. Quality evaluation of the decoction of L. chuanxiong Hort

The content of the representative active ingredients, the solid content, and the HPLC fingerprints similarity were used as the quality monitoring indicators of *L. chuanxiong* decoction before and after the removal of heavy metals.

## 2.10.1. Determination of Senkyunolide I, Senkyunolide A, and Ferulic acid

Senkyunolide I, Senkyunolide A, and Ferulic acid were chosen as representative active ingredients of the decoction of *L*. *chuanxiong* Hort. These compounds were analyzed and compared in the decoction before and after removing heavy metals based on HPLC analysis. The HPLC parameters were described in section 2.2.

#### 2.10.2. Analysis of the solid content

25 mL of *L. chuanxiong* Hort decoction before and after treatment were respectively transferred to an evaporation dish that was dried to constant weight and evaporated to dryness in a boiling water bath. Then, the solid content was weighed and measured.

#### 2.10.3. Evaluation of HPLC fingerprints similarity

The *L. chuanxiong* Hort decoction was analyzed by HPLC before and after the removal of heavy metals. Then, the similarity of HPLC fingerprints was evaluated by the similarity evaluation system for the chromatographic fingerprints of traditional Chinese medicine (2012 Edition).

#### 3. Results and discussion

#### 3.1. Characterization of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG

The FT-IR spectrum of the in-house fabricated material is shown in Fig. S1. It exhibited the FT-IR spectra of  $Fe_3O_4$ ,  $Fe_3-O_4$ @PDA,  $Fe_3O_4$ @MIL-101(Cr),  $Fe_3O_4$ @MIL-101(Cr) @PEG, and MIL-101(Cr). In the spectrum of  $Fe_3O_4$ , the characteristic absorption peak at 606 cm<sup>-1</sup> was attributed to the Fe-O-Fe vibration. The two broad bands at 3457 cm<sup>-1</sup> and 1640 cm<sup>-1</sup> were respectively assigned to O–H stretching and bending vibrations of water, which were probably located on the surface of iron oxide nanoparticles(Shahabadi et al., 2016). In addition to the typical peaks of  $Fe_3O_4$ , many new peaks can be seen in the spectrum of  $Fe_3O_4$  (*i*)PDA. Representatively, the peak at 630 cm<sup>-1</sup> was ascribed to the N-H bending vibration. The unique absorption bands near 1290 and 1500 cm<sup>-1</sup> were related to the bending vibration of C-N and the stretching vibration of the aromatic ring, respectively. These features showed that the PDA had been successfully coated on the surface of Fe<sub>3</sub>O<sub>4</sub>. As for MIL-101(Cr), the band at 1021 cm<sup>-1</sup> was corresponded to the in-plane bend vibrations of terephthalic acid. The deformation vibration belonging to C-H at 1150, 1013, 889, 751 cm<sup>-1</sup> was consistent with the reported peaks of MIL-101(Cr) in literature(Oiu et al., 2020). And as for  $Fe_3O_4$  (@MIL-101(Cr), the characteristic vibration bands of O-C-O (around 1385 and 1624 cm<sup>-1</sup>) indicated the existence of the terephthalic acid within MIL-101(Cr) and the connection of the carboxylate ligand to the Cr metal center, which could be obtained in MIL-101(Cr) spectrum (Yalcin & Kayan, 2012). The broad band near 585 cm<sup>-1</sup> confirmed the incorporation of the Fe-O-Fe group in the nanocomposite(Madhuri et al., 2005). As for Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG, the shifting and increased intensity of the peak in the region  $3500-3300 \text{ cm}^{-1}$  indicated the enhanced hydrogen bonding induced by the hydroxyl groups of PEG. The results demonstrated the successful fabrication of Fe<sub>3</sub>O<sub>4</sub>@-MIL-101(Cr)@PEG.

The morphology of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) and Fe<sub>3</sub>O<sub>4</sub>@-MIL-101(Cr)@PEG was investigated by SEM and TEM. As described in our last work, the MIL-101(Cr) was verified to be randomly extended around Fe<sub>3</sub>O<sub>4</sub>@PDA (Fig. S2a and Fig. S2c)(Tang et al., 2022). In comparison to Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr), a rougher surface was obviously observed for Fe<sub>3</sub>-O<sub>4</sub>@MIL-101(Cr)@PEG, which was speculated to be the extension of PEG around the Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) (Fig. S2b). The TEM image of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG further verified that PEG was loaded around Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) to construct a core-shell structure (Fig. S2d). The results showed that a larger specific surface area was obtained after the modification of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) by PEG. Besides, the overlapped and agglomerated PEG provided a restricted-access structure, which was beneficial for specifically trapping heavy metals.

Since the specific surface area is another important property of adsorbent material, the porous structure and property of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG were tested in this study. Generally, type IV isotherm is generated by mesoporous solids. A typical feature is that the adsorption branch of the isotherm is inconsistent with the desorption branch of the isotherm, and a hysteresis loop can be observed. In this study, a type IV curve with a small hysteresis was provided as the N2 adsorption-desorption isotherm of the material in Fig. S3, indicating the typical property of a mesoporous structure. The Brunauer-Emmett-Teller (BET) surface area was 248.5447 m<sup>2</sup>/g, which was satisfactory for enrichment purposes. Moreover, the total pore volume was calculated to be  $0.17 \text{ cm}^3/\text{g}$  by the Barrett-Joyner-Halenda method. The average pore diameter of the amorphous nanomaterial was about 4.59 nm. The results have verified that the in-house fabricated Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) @PEG had a high surface area, large pore volume, and narrow pore diameter. According to the above results, the Fe<sub>3</sub>O<sub>4</sub>(a)-MIL-101(Cr)@PEG was characterized by beneficial structure

and properties for adsorption and separation of heavy metals, which was simultaneously expected to avoid the influence on the active ingredients.

XRD patterns were used to identify the crystal structure of the synthesized nanomaterials(Xiong, Liu, et al., 2023). Fig. S4 exhibited the XRD patterns of MIL-101(Cr) and Fe<sub>3</sub>O<sub>4</sub>@-MIL-101(Cr)@PEG. The sharp diffraction peaks indicated that the MIL-101(Cr) was a material with high crystallinity. The obvious peaks at 20 of 5.0°, 5.9°, 9.8°, 10.5°, represented the reflections of typical MIL-101(Cr)(Li et al., 2020). After intercalating Fe<sub>3</sub>O<sub>4</sub> nanoparticles into MIL-101(Cr), the characteristic peaks of Fe<sub>3</sub>O<sub>4</sub> at 30.0°, 35.4°, 54.5°, 63.2°, 67.6° could be observed(Davodi et al., 2019). Obviously, the characteristic diffraction peaks of MIL-101(Cr) were presented in the Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG, indicating that the Fe<sub>3</sub>O<sub>4</sub> nanoparticles didn't change the high crystallinity of MIL-101 (Cr)(Abutalib & A.Rajeh, 2020). In addition, the peak at 20 of 19.1° was the characteristic peak of PEG(Zheng et al., 2021), which coincided with the peak of MIL-101(Cr). But it could be confirmed from the intensity of the peak that it was the characteristic peak of PEG. All these characterizations supported the successful preparation of the Fe<sub>3</sub>O<sub>4</sub>@MIL-101 (Cr)@PEG nanocomposites.

The magnetic properties of the Fe<sub>3</sub>O<sub>4</sub> and Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG were studied using a VSM in an external magnetic field from - 12,000 Gs to + 12,000 Gs. As shown in Fig. S5, the hysteresis loops of Fe<sub>3</sub>O<sub>4</sub> and Fe<sub>3</sub>O<sub>4</sub>@MIL-101 (Cr)@PEG were no hysteresis, which showed no remanence or coercivity at room temperature. The results exhibited the typical superparamagnetic character of both materials. The maximal saturation magnetization of Fe<sub>3</sub>O<sub>4</sub> was 62.98 emu/ g. As for Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG, the saturation magnetization decreased to 40.42 emu/g. Although the saturation magnetization was lower than that of bulk magnetite (92 emu/g)(Han et al., 1994), the particles were only redispersed under slight shaking when the magnetic field was removed. This feature facilitated collection, regeneration, and reutilization of the materials. Thus, it could ensure the easy separation of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG from the aqueous solution with a low magnetic field. These structural characteristics and magnetic behavior make Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) @PEG nanocomposites great potential as a good adsorbent for removing heavy metals from Chinese medicinal materials.

### 3.2. Removal of heavy metals in the decoction of L. chuanxiong Hort

Parameters that can influence the removal efficiency of the heavy metal ions were initially investigated. All experiments were conducted in triplicate with 50 mL of spiked *L. chuanxiong* Hort decoction, of which the pH was adjusted with NaOH solution. The spiked concentration of each heavy metal ion was 50  $\mu$ g/mL, including Pb, Cd, As, Ni, and Zn ions. At the same time, parallel experiments were conducted on *L. chuanxiong* Hort decoction without spiking heavy metals for the deduction of background.

#### 3.2.1. Effect of pH

The solution pH is a key factor for the adsorption removal of heavy metal ions since it could affect the existing forms of heavy metals and the surface charge of the adsorbents by changing the protonation degree of its functional groups. Therefore, the influence of solution pH on the removal efficiency of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG toward Pb<sup>2+</sup>, Cd<sup>2+</sup>, As<sup>5+</sup>, Zn<sup>2+</sup>, and Ni<sup>2+</sup> is worth studying. In optimization of pH (pH 2-pH 8), the adsorption time and the adsorbent amount were 60 min and 25 mg, respectively. As shown in Fig. 2A, Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG exhibited better adsorption performance toward Pb<sup>2+</sup>, Cd<sup>2+</sup>, and As<sup>5+</sup> in a suitable pH range. It displayed that the removal efficiency of  $Fe_3O_4(a)$ -MIL-101(Cr)@PEG for all heavy metal ions was lower at low pH and enhanced with increasing pH within a certain range. At the highly acidic solutions, the amine groups and oxygen atoms with lone electrons in the material would be protonated due to the existence of many hydrogen ions. Thus, the positively charged groups on the surface of the adsorbents may repulse the positively charged metal ions. Moreover,  $H_3O^+$ ions competed with other adsorbate ions for the vacant binding sites on the adsorbents in acidic media. As a consequence, the removal efficiency of all metal ions decreased in conditions of low pH(Kayan & Kayan, 2022). Although the surface charge of the adsorbents may be negative at higher pH, the interactions seemed to be reduced between heavy metal ions and the adsorbents, which might be ascribed to the charge neutralization between the heavy metal and hydroxyl ions. In summary, the removal efficiency increased from pH 2 to pH 6 while slightly decreased from pH 6 to pH 8. It is worth mentioning that the actual pH value of *L. chuanxiong* Hort decoction is 5.36. The removal efficiency of the heavy metal ions was additionally tested at pH 5.36. As can be seen from the result, the best removal efficiency was obtained at pH 5.36 for Pb<sup>2+</sup>, while pH 6 for As<sup>5+</sup> and Ni<sup>2+</sup>. Besides, the removal efficiency was almost the same for Zn<sup>2+</sup> and Cd<sup>2+</sup> at two pH values. Comprehensively considering the removal efficiency and the convenience of practical application, the original pH of the decoction (pH 5.36) was applied in the following experiment.

#### 3.2.2. Effect of adsorption time

The adsorption time was evaluated in the range of 30–360 min. In optimization of adsorption time, the solution pH and the adsorbent amount were applied as 5.36 and 25 mg, respec-



Fig. 2 Effect of (A) solution pH, (B) adsorption time, and (C) adsorbent amount on the removal efficiency of heavy metals.

tively. Fig. 2B displayed the effect of the adsorption time on the removal efficiency of As, Ni, Zn, Cd, and Pb ions. It can be found that the removal rate was proportional to the adsorption time in a certain adsorption time. The removal efficiency of  $Zn^{2+}$ ,  $Cd^{2+}$ , and  $Pb^{2+}$  almost remained the same from the adsorption time of 240 min, and the removal efficiency of Ni<sup>2+</sup> and As<sup>5+</sup> reached the equilibrium at the adsorption time of 300 min. Specifically, the adsorption rates followed



Fig. 3 The adsorption kinetics of  $As^{5+}$ ,  $Ni^{2+}$ ,  $Zn^{2+}$ ,  $Cd^{2+}$ , and  $Pb^{2+}$  onto  $Fe_3O_4@MIL-101(Cr)@PEG$  (A) and fitted lines with pseudo-first-order (B) and pseudo-second-order (C).

Table 1Kinetic factors of Fe <sub>3</sub> O <sub>4</sub> @MIL-101(Cr)@PEG toward five heavy metal ions.							
Kinetic models	Metal ions	$\mathbb{R}^2$	Qe (mg/g)	K			
Pseudo-first-order	As Ni Zn	0.9835 0.9965 0.9802	$\begin{array}{r} 45.90 \ \pm \ 1.33 \\ 26.33 \ \pm \ 0.37 \\ 28.27 \ \pm \ 0.75 \end{array}$	$\begin{array}{r} 0.0109 \ \pm \ 0.0010 \\ 0.0108 \ \pm \ 0.0005 \\ 0.0128 \ \pm \ 0.0012 \end{array}$			
Pseudo-second-order	Pb Cd As	0.9808 0.9883 0.9974	$\begin{array}{rrrr} 49.67 \ \pm \ 1.20 \\ 46.37 \ \pm \ 0.96 \\ 57.83 \ \pm \ 1.07 \end{array}$	$\begin{array}{r} 0.0139 \ \pm \ 0.0012 \\ 0.0128 \ \pm \ 0.0009 \\ 0.0002 \ \pm \ 0.00001 \end{array}$			
	Ni Zn Pb Cd	0.9954 0.9978 0.9989 0.9948	$\begin{array}{r} 33.45 \pm 0.87 \\ 34.61 \pm 0.49 \\ 59.91 \pm 0.55 \\ 56.83 \pm 0.33 \end{array}$	$\begin{array}{r} 0.0003 \ \pm \ 0.00003 \\ 0.0004 \ \pm \ 0.00002 \\ 0.0003 \ \pm \ 0.00001 \\ 0.0002 \ \pm \ 0.00005 \end{array}$			



Fig. 4 The removal efficiency of  $Fe_3O_4$ , MIL-101(Cr), and  $Fe_3-O_4$ @MIL-101(Cr)@PEG for five heavy metal ions.

the sequence:  $Cd^{2+} \approx Pb^{2+} > As^{5+} > Zn^{2+} > Ni^{2+}$ , which could be due to the difference in the hydrated radius, electronegativity, and acid-base properties. To sum up, the best adsorption time was 240 min.

Furthermore, based on Fig. 3 and the data summarized in Table 1, it was evident that the pseudo-second-order kinetic model was more suitable for describing the heavy metals adsorption on Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG, suggesting that the adsorption of heavy metals on the Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) @PEG mainly occurred through chemical adsorption(Xue Zhang et al., 2020). The chemical adsorption probably involved the electrons exchange and sharing (-OH, -NH<sub>3</sub>, strong chelating groups), and covalent interactions of active sites with metal ions (As<sup>5+</sup>, Ni<sup>2+</sup>, Zn<sup>2+</sup>, Cd<sup>2+</sup>, and Pb<sup>2+</sup>) (Xiong et al., 2020; Xiong et al., 2021; Xue Zhang et al., 2020).

#### 3.2.3. Effect of adsorbent amount

The adsorbent amount is another crucial parameter affecting the removal efficiency of heavy metals. The effect of the adsorbent amount was investigated in the range from 10 mg to 60 mg. In optimization of adsorbent amount, the solution pH and the adsorption time were 5.36 and 240 min, respectively. Generally, the ratio of adsorbate ions to adsorbent



Fig. 5 (A) The adsorption isotherms of  $As^{5+}$ ,  $Ni^{2+}$ ,  $Zn^{2+}$ ,  $Cd^{2+}$ , and  $Pb^{2+}$  onto  $Fe_3O_4@MIL-101(Cr)@PEG$ , (B) fitted lines with the Langmuir model, and (C) fitted lines with the Freundlich model.

Table 2 Isotherm parameters for the adsorption removal of five heavy metal ions on Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG.

Isotherm models	Constants	Heavy metal ions							
		As	Ni	Zn	Pb	Cd			
Langmuir	$\begin{array}{c} Q_m \ (mg/g) \\ K_L \ (L/mg) \\ R^2 \\ R_L \end{array}$	$\begin{array}{r} 60.27 \ \pm \ 2.68 \\ 0.1656 \ \pm \ 0.0277 \\ 0.9872 \\ 0.057 - 0.883 \end{array}$	$38.48 \pm 1.58 \\ 0.0758 \pm 0.0103 \\ 0.9924 \\ 0.1166 - 0.9428$	$\begin{array}{r} 45.06 \ \pm \ 1.71 \\ 0.0818 \ \pm \ 0.0103 \\ 0.9935 \\ 0.1089 - 0.9386 \end{array}$	$72.18 \pm 3.28 \\ 0.0979 \pm 0.0142 \\ 0.9923 \\ 0.0927-0.9274$	$\begin{array}{r} 62.29 \pm 0.91 \\ 0.1442 \pm 0.0079 \\ 0.9988 \\ 0.0649 \\ 0.8966 \end{array}$			
Freundlich	$ \begin{array}{c} K_{\rm F} \ ({\rm mg/g}) \\ 1/n_{\rm F} \\ {\rm R}^2 \end{array} $	$\begin{array}{l} 11.8986 \pm 3.0280 \\ 0.3940 \pm 0.0716 \\ 0.9119 \end{array}$	$\begin{array}{l} 4.9135 \pm 1.2962 \\ 0.4559 \pm 0.0686 \\ 0.9413 \end{array}$	$5.9556 \pm 1.3472 \\ 0.4558 \pm 0.0597 \\ 0.9565$	$\begin{array}{r} 9.8464 \pm 2.2955 \\ 0.4763 \pm 0.0664 \\ 0.9494 \end{array}$	$\begin{array}{c} 11.2755 \pm 2.3451 \\ 0.4155 \pm 0.0587 \\ 0.9519 \end{array}$			

Table 3 Comparison of the maximum adsorption capacity of different adsorbents toward five heavy metal ions.

Adsorbent	Adsorption ca	pacity (mg/g)	Reference			
	As <sup>5+</sup>	Ni <sup>2+</sup>	$Zn^{2+}$	$Pb^{2+}$	$Cd^{2+}$	
UiO-66-NHC(S)NHMe	-	_	-	232	49	(Saleem et al., 2016)
HFB	_	12.13	22.25	2.89	22.03	(Shin, 2017)
Denim textile wastes	1.50	_	25.50	-	14.83	(Mendoza-Castillo et al., 2015)
Granular activated carbon	2.50	-	_	_	_	(Natale et al., 2008)
Fe <sub>3</sub> O <sub>4</sub> /Mg-Al-CO <sub>3</sub> -	-	-	-	-	54.7	(Shan et al., 2015)
LDH						
Rice husk ash	-	_	14.30	-	-	(Ahmaruzzaman, 2011)
LDH-Cl	-	_	_	108	61	(González et al., 2015)
Modified waste silk	_	7.47	_	_	8.03	(Mia et al., 2022)
Chitosan	_	-	_	7.70	_	(Cheng & Lin, 2006)
Fe <sup>2+</sup> -loaded activated carbon	2.02	-	_	_	_	(Tuna et al., 2013)
Iron-modified carbon	1.25	-	-	_	_	(Vitela-Rodriguez & Rangel- Mendez, 2013)
Dashukivskij bentonite	_	-	_	52.08	26.04	(Kostenko et al., 2019)
Poultry litter	_	52	80	_	-	(Kucharski et al., 2021)
MgAL-LDH	_	-	_	12	18.25	(Xue Zhang et al., 2020)
Hybrid material-A	_	_	14.29	38.76	_	(Cerraholu et al., 2018)
Hybrid material-B	_	_	25.61	38.09	-	(Cerraholu et al., 2018)
Fe <sub>3</sub> O <sub>4</sub> @MIL-101(Cr) @PEG	$60.27~\pm~2.68$	38.48 ± 1.58	45.06 ± 1.71	$72.18~\pm~3.28$	$62.29~\pm~0.91$	This work

vacancies was reduced by increasing the adsorbent dosage. Thus, the removal efficiency increased with the adsorbent amount, as shown in Fig. 2C. However, at higher adsorbent dosages, the available vacancies may be reduced due to the aggregation of the adsorbent nanoparticles. The removal efficiency was accordingly predicted to decrease. In this study, the removal efficiency of  $As^{5+}$ ,  $Ni^{2+}$ ,  $Zn^{2+}$ ,  $Cd^{2+}$ , and  $Pb^{2+}$  were close to the maximum when the adsorbent amount was 40 mg. With the adsorbent amount continually increasing from 40 mg to 60 mg, the removal efficiency of heavy metals remained basically unchanged.

In summary, the optimized parameters were obtained as follows: pH 5.36, adsorption time of 240 min, and adsorbent amount of 40 mg. Under the optimal conditions and the spiked concentration of 50 µg/mL, the removal efficiency of Fe<sub>3</sub>O<sub>4</sub>@-MIL-101(Cr)@PEG for As<sup>5+</sup>, Ni<sup>2+</sup>, Zn<sup>2+</sup>, Cd<sup>2+</sup>, and Pb<sup>2+</sup> were 81.40 %, 51.59 %, 55.40 %, 88.19 %, and 83.79 %, respectively. It indicated that the in-house fabricated adsorbent had excellent removal efficiency for As<sup>5+</sup>, Cd<sup>2+</sup>, and Pb<sup>2+</sup> in the decoction of *L. chuanxiong* Hort, while obvious removal efficiency for Ni<sup>2+</sup> and Zn<sup>2+</sup>.

To further assess the removal efficiency of the Fe<sub>3</sub>O<sub>4</sub>@-MIL-101(Cr)@PEG, comparison with Fe<sub>3</sub>O<sub>4</sub> and MIL-101 (Cr) was carried out under the same conditions. As can be seen in Fig. 4, the Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG exhibited the highest removal efficiency for As<sup>5+</sup>, Ni<sup>2+</sup>, Zn<sup>2+</sup>, Cd<sup>2+</sup>, and Pb<sup>2+</sup>.

#### 3.2.4. Adsorption isotherms

Adsorption isotherms are critical for the overall improvement of adsorption mechanism pathways and the effective design of an adsorption system. The isotherm data are analyzed using the Langmuir and Freundlich models. The relevant results were depicted in Fig. 5, and the measured parameters were compiled in Table 2. As presented in Fig. 5A, the adsorption capacity of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG toward As<sup>5+</sup>, Ni<sup>2+</sup>, Zn<sup>2+</sup>, Cd<sup>2+</sup>, and Pb<sup>2+</sup> increased rapidly with an increase in their initial concentrations till the plots reach a plateau and remained relatively constant. The emergence of the plateau signifies that active sites were fully saturated(Ahmadijokani et al., 2021). Fig. 5 unfolded that the Langmuir isotherm model matched the experimental results.

Temperature (°C)	Constants	Heavy metal ions						
		As	Ni	Zn	Pb	Cd		
20	$\Delta G(kJ \cdot mol^{-1})$	-24.33	-7.11	-8.98	-17.22	-21.89		
	$\Delta H(kJ \cdot mol^{-1})$	-17.65	-2.95	-3.84	-6.02	-14.12		
	$\Delta S(J \cdot mol^{-1} \cdot K^{-1})$	-41.07	-1.18	-2.27	-4.27	-29.93		
	$Q_m (mg/g)$	$60.27~\pm~2.68$	$38.48~\pm~1.58$	$45.06 \pm 1.71$	$72.18~\pm~3.28$	$62.29~\pm~0.91$		
	$\mathbb{R}^2$	-	-	-	-	-		
30	$\Delta G(kJ \cdot mol^{-1})$	-19.75	-7.04	-8.81	-16.43	-18.60		
	$\Delta H(kJ \cdot mol^{-1})$	-	-	-	-	-		
	$\Delta S(J \cdot mol^{-1} \cdot K^{-1})$	-	_	-	-	_		
	$Q_m (mg/g)$	$55.02 \pm 1.34$	$27.77 \pm 0.83$	$34.76 \pm 1.80$	$62.92 \pm 3.10$	$55.82~\pm~2.88$		
	$R^2$	0.9998	0.9949	0.9995	0.9999	0.9992		
40	$\Delta G(kJ \cdot mol^{-1})$	-16.36	-7.03	-8.68	-15.71	-16.15		
	$\Delta H(kJ \cdot mol^{-1})$	-	-	_	-	_		
	$\Delta S(J \cdot mol^{-1} \cdot K^{-1})$	-	-	-	-	-		
	$Q_m (mg/g)$	$47.18 \pm 1.21$	$19.62 \pm 0.71$	$26.92 \pm 1.35$	$54.32 \pm 2.92$	$50.19 \pm 1.05$		
	$\mathbb{R}^2$	-	-	-	-	-		

**Table 4** Thermodynamic parameters for the adsorption removal of five heavy metal ions on  $Fe_3O_4@MIL-101(Cr)@PEG$  at different temperatures.



Fig. 6 Reusability of the in-house fabricated  $Fe_3O_4@MIL-101$  (Cr)@PEG.

Same conclusion was presented in Table 2. It implied that the adsorption of  $As^{5+}$ ,  $Ni^{2+}$ ,  $Zn^{2+}$ ,  $Cd^{2+}$ , and  $Pb^{2+}$  onto the Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG was monolayer. There was no interaction between adsorbate ions, and the surface of the adsorbent was homogeneous(Yuan et al., 2017). As can be seen from Table 2, the calculated  $R_L$  and  $1/n_F$  were between 0 and 1, indicating that the adsorption process was favorable and Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG was a good adsorbent for heavy metals removal from aqueous solution. This result suggested that the adsorption process of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) @PEG toward As<sup>5+</sup>, Ni<sup>2+</sup>, Zn<sup>2+</sup>, Cd<sup>2+</sup>, and Pb<sup>2+</sup> was monolayer coverage adsorption and favorable(Mahmoodi & Mohammad, 2015; Xue Zhang et al., 2020).

Moreover, the maximum adsorption capacity of the adsorbent toward the heavy metals was compared with the literature data in Table 3. The adsorption capacity of the five heavy metal ions on  $Fe_3O_4@MIL-101(Cr)@PEG$  was equal to or better than other adsorbents. Some data showed that the adsorption capacity of this material for heavy metal ions was



**Fig. 7** XRD patterns of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG before and after being used in different conditions.

lower than that of other materials, and it was because those other materials only target one or two heavy metal ions, which greatly reduced the competition for material adsorption sites.

The temperature of an adsorption system is associated with adsorption thermodynamics and affects the adsorption efficiency. Thermodynamic adsorption parameters were measured and shown in Table 4. It can be seen that the theoretical maximum adsorption amount ( $Q_m$ ) of five heavy metals on Fe<sub>3</sub>-O<sub>4</sub>@MIL-101(Cr)@PEG decreased in turn with an increase in temperature from 20 °C to 40 °C, indicating that lowering the temperature was beneficial to the adsorption process. This phenomenon can be explained by the negative values of  $\Delta H$  of  $As^{5+}$ ,  $Ni^{2+}$ ,  $Zn^{2+}$ ,  $Cd^{2+}$ , and  $Pb^{2+}$ , which implied that the adsorption process was exothermic. Hence, lowering the temperature was conducive to adsorption. As for the  $\Delta G$  values all less than zero at different temperatures, it indicated that



Fig. 8 HPLC chromatogram of (A) the mixed standard and (B) the testing sample.

the adsorption process was spontaneous. In addition, the negative values of  $\Delta S$  indicated that the degree of freedom of the interface between Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG and the aqueous solution decreased during the adsorption process. To sum up, the adsorption of As<sup>5+</sup>, Ni<sup>2+</sup>, Zn<sup>2+</sup>, Cd<sup>2+</sup>, and Pb<sup>2+</sup> by Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG was a spontaneous, exothermic, and entropy decreasing process.

#### 3.3. Adsorbent reusability

The reusability of the adsorbent was investigated by adsorption–desorption cycle experiments, which was an important parameter to evaluate the applicability of Fe<sub>3</sub>O<sub>4</sub>@MIL-101 (Cr)@PEG. As shown in Fig. 6, the removal efficiency for As, Ni, Zn, Pb, and Cd ions almost remained unchanged after four consecutive adsorption–desorption cycles. The relative standard deviations (RSDs) were calculated to be 5.47 % (As<sup>5+</sup>),

8.27 % (Ni<sup>2+</sup>), 5.51 % (Zn<sup>2+</sup>), 7.45 % (Cd<sup>2+</sup>), and 6.60 % (Pb<sup>2+</sup>), respectively. It was demonstrated that the removal efficiency of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG for As, Ni, Zn, Pb, and Cd ions had been well restored at least for four cycles. The result indicated that the in-house fabricated Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) @PEG had good reusability and a great potential for sustainable heavy metals removal. In addition, the amount of material after the first heavy metal desorption was 34.4 mg. The loss rate was only 14 %, which meant that the recovery rate of the material was 86 %. In this way, only a small amount of adsorbents was added in successive experiment. In general, the recovery rate of the adsorbent was satisfactory.

#### 3.4. Adsorbent stability

The XRD patterns of Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG were shown in Fig. 7. After being used in different solutions (HCl

Table 5 Content of the active ingredients in the decoction of *L. chuanxiong* Hort before and after the heavy metal removal.

Compounds	Quality score		Rate of change (%)				
	Control	Sample 1	Sample 2	Sample 3	Sample 4	Sample 5	
Ferulic acid	0.1300	0.1189	0.1216	0.1211	0.1211	0.1207	7.18
Senkyunolide I	0.0459	0.0443	0.0456	0.0448	0.0459	0.0447	1.93
Senkyunolide A	0.1283	0.1213	0.1245	0.1234	0.1240	0.1223	4.09



Fig. 9 Similarity evaluation on HPLC fingerprints of L. chuanxiong decoction with  $(A_{1-5})$  and without (A) the heavy metals removal.

and heavy metals solution) for 10 h, the changes in the crystalline structure of the MOF were insignificant. It was revealed that the Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG was quite stable in the acidic solution. Moreover, the material was stable after absorbing heavy metal ions. The obtained results indicated that the synthesized adsorbent had good structural stability and was suitable for the removal of As, Ni, Zn, Cd, and Pb ions from the *L. chuanxiong* Hort decoction.

#### 3.5. Metal leaching rate

The metal ion leaching of the material was tested after use. The resulted leaching rate of Fe ion was 0.00016 % and that of Cr ion was 0.0049 %, indicating that Fe and Cr ions would not cause secondary pollution to *L. chuanxiong* Hort decoction.

# 3.6. Influence on L. chuanxiong Hort decoction after heavy metals removal

The quality monitoring indicators of *L. chuanxiong* Hort decoction were evaluated before and after the removal of heavy metals. In this section, the experiments of heavy metals

removal were directly carried out on the *L. chuanxiong* Hort decoction.

### 3.6.1. The content of Senkyunolide I, Senkyunolide A, and Ferulic acid

Since Ferulic acid and Senkyunolide are the main effective components of *L. chuanxiong* Hort, of which the pharmacological effects have been well studied(Lanlan et al., 2022; Zhen et al., 2011), Senkyunolide A, Senkyunolide I, and Ferulic acid were selected as the indicators to evaluate the influence of the adsorbent on the effective components. To evaluate the influence of the adsorption process on the active ingredients of *L. chuanxiong* Hort decoction, Senkyunolide I, Senkyunolide A, and Ferulic acid were analyzed in the same decoction with and without the removal of heavy metals. The HPLC chromatogram of Senkyunolide I, Senkyunolide A, and Ferulic acid was shown in Fig. 8.

The content of Senkyunolide I, Senkyunolide A, and Ferulic acid in the decoction of *L. chuanxiong* Hort before and after the heavy metals removal was listed in Table 5. Because the change rate was less than 8.00 %, it was clear that the content of the active ingredients was not significantly affected by the adsorption removal process of heavy metals. The in-house fabricated adsorbent exhibited to be able to efficiently remove heavy metals in *L. chuanxiong* Hort decoction without influencing its active ingredients.

#### 3.6.2. Solid content of the decoction

The average solid content of the solution before and after heavy metals removal was  $18.19 \text{ mg} \cdot \text{mL}^{-1}$  and  $18.16 \text{ mg} \cdot \text{mL}^{-1}$ , respectively. The loss rate was only 0.16 %, indicating no obvious loss of solid content before and after removing heavy metals with the in-house fabricated adsorbent.

#### 3.6.3. HPLC fingerprints similarity

With the same decoction, one sample solution was directly analyzed by HPLC to acquire the chromatographic fingerprint, while the other five were analyzed after the removal of heavy metals. The similarity evaluation values were all calculated to be greater than 99.9 %. Thus, there was no significant change in the number and area of the chromatographic peaks of *L. chuanxiong* decoction before and after the removal of heavy metals (Fig. 9). The consistency of the chromatogram profiles indicated that the chemical composition of the decoction remained almost complete before and after treatment. The results further verified that the in-house fabricated adsorbent wound not affect the chemical components of *L. chuanxiong* decoction during the adsorption removal process of heavy metals.

#### 4. Conclusion

In this study, a new magnetic nanomaterial (Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr) @PEG) was successfully synthesized to remove heavy metals based on the PDA bridging strategy and PEG functionalization (PEG played a role in blocking the adsorption of effective components of L. chuanxiong Hort). The introduction of magnetism in the material allowed us to simply separate the adsorbent from the decoction of L. chuanxiong Hort by using an external magnetic field. The in-house fabricated adsorbent exhibited strong magnetic responsiveness, excellent hydrophilicity, and good recyclability. The kinetic and isotherm adsorption studies showed that the adsorption of these metal ions onto Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG could be described well by pseudosecond-order kinetic model as well as the Langmuir isotherm model, suggesting that the adsorption is chemisorption and monolayer. The maximum adsorption capacity of the as-prepared Fe<sub>3</sub>O<sub>4</sub>@MIL-101 (Cr)@PEG reached 60.27 mg/g ( $As^{5+}$ ), 38.48 mg/g ( $Ni^{2+}$ ), 45.06 mg/g ( $Zn^{2+}$ ), 72.18 mg/g ( $Pb^{2+}$ ), and 62.29 mg/g ( $Cd^{2+}$ ). Thermodynamic studies revealed that the adsorption of heavy metal ions on Fe<sub>3</sub>O<sub>4</sub>@MIL-101(Cr)@PEG was spontaneous and exothermic within the experimental temperatures. Based on these properties, the adsorbent was verified to have good performance in the removal of Pb<sup>2+</sup>, Cd<sup>2+</sup>, As<sup>5+</sup>, Ni<sup>2+</sup>, and Zn<sup>2+</sup> from *L. chuanxiong* decoction. In a short adsorption time, removal efficiency in the range of 51.59 %-88.19 % was obtained for the spiked heavy metals (50 µg/mL) without any significant influence on the quality of L. chuanxiong decoction, which was represented by the content of active ingredients, solid content, and chromatogram profile. In summary, the proposed adsorbent provided considerable removal efficiency for  $As^{5+}$ ,  $Cd^{2+}$ , and  $Pb^{2+}$  and certain removal efficiency for  $Ni^{2+}$  and  $Zn^{2+}$ , exhibiting great potential for heavy metals removal in medicinal herbs. Since only one kind of medicinal herb was considered, further studies are needed concerning more species of medicinal herbs and different functionalization strategies of the adsorbent.

#### **Declaration of Competing Interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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#### Appendix A. Supplementary material

Supplementary data to this article can be found online at https://doi.org/10.1016/j.arabjc.2023.104635.

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