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ORIGINAL ARTICLE

Simultaneous removal and conversion of silver ions from wastewater into antibacterial material through selective chemical precipitation



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KEYWORDS

Silver recovery; Wastewater; Selective chemical precipitation; Antibacterial material; Planar precipitant **Abstract** High doses of silver compounds in water have been found to be extremely toxic to the cells of plants and animals. Removal then reutilization of silver ions as antibacterial materials in one step from wastewater is an urgent but challenging issue from the views of environment and economy. In this research, we reported a planar benzothiazole Schiff base derivative **BTS1** that can selectively precipitate silver ions from wastewater over other metal ions through chemical precipitation. The precipitation reaction showed promising activity in the pH range of 6 to11, realizing a high silver ions precipitation ratio of 95.4%, and low LOD of 0.059 mg/L. More importantly, the filtrated precipitate **BTS1-Ag** can be directly used as an antibacterial material. The test papers coated with **BTS1-Ag** exhibited antibacterial activity against both gram-positive and gram-negative bacteria. This research provides a promising strategy for simultaneous removal and conversion of silver ions from wastewater into functional materials.

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1. Introduction

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The removal and recovery of precious metals from wastewater have drawn great attention due to the low abundance of these raw metal resources and risk in ecology. (Ling et al., 2018; Qian et al., 2020; Yang et al., 2020; Wang et al., 2023) As one of the most important and useful precious metals, silver possesses wide application in electroplating, photography, batteries, electronics, (Ibrahim et al., 2022) catalysis, (Xue et al., 2020) biomedical, (Chernousova and Epple, 2013) photonics and dental materials due to its unique photosensitivity and bioactivity. (Wang et al., 2020) In this context, a large amount of silver ions-rich industrial wastewater has been discharged into the

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natural environment. Recently, growing concerns have been raised on its accumulated toxicity on the cells of plants and animals. (Ratte, 1999; Ahmad et al., 2019; Xie et al., 2022) Besides, direct discharge of silver ions-rich wastewater is also a waste of resources in consideration of the limited stock of precious metals on earth. (Akcil et al., 2015; Sverdrup et al., 2014; Wang et al., 2020) In 2011, the US Water Environment Research Foundation (WERF) reported that treating wastewater 10 million gallon per day may have the potential to recover \$8849-\$3,904,664 worth of silver worth per year. (Ho and Babel, 2021) Therefore, the highly selective recovery of silver ions from the wastewater containing multi-metal ions for further utilization demonstrates extreme importance from the view of environment and economy. To date, plenty of methods and techniques including membrane filtration, electrolysis, chemical precipitation, cementation, adsorption, ion exchange, electrowinning, solvent extraction and bioelectrochemical system have been developed to recover silver from wastewater. (Wang and Ren, 2014; Xu et al., 2018; Wen et al., 2013) The chemical precipitation attracts wide attentions on account of its eco-friendship, low cost, and easy operation. (Fu et al., 2007; Ying and Fang, 2006; Chen et al., 2018; Quiton et al., 2022).

As the core of chemical precipitation, various inorganic sulfide precipitants have been adopted for silver recovery, such as Na₂S, K₂S, NaHS, NH4HS and Na2S2O3/Na2S2O4 system. In these cases, strict dosage control is needed to avoid possible accidental escape of the toxic hydrogen sulfide. Additionally, the poor selectivity also goes against the exclusive precipitation of silver ions over other metal ions, such as Ca^{2+} , Cu^{2+} and Pb^{2+} . (Bas et al., 2012; Yazici et al., 2017) More importantly, the separated silver sulfide requires further refining process to obtain purified silver or combining with other materials to acquire new functions, which undoubtedly consumes additional energy and resource. (Ho and Babel, 2021) An ideal technology for the treatment of wastewater containing precious metal ions should be able to achieve two goals simultaneously: targeted precious metal ions removal and conversion into useful materials directly. (Ling et al., 2018, W.X. Zhang, 2018) Recently, the prevalence of Covid-19 has forced researchers spanning the globe to focus on the safe and effective antibacterial agents. In this context, silver nanoparticles and silver (I) (Ag⁺) complexes with antibacterial activities have received rising attention. (Hamouda et al., 2021; Allawadhi et al., 2021; Morozova et al., 2022; Law et al., 2022; Teirumnieks et al., 2020; Abbas et al., 2022) Thereafter, it is of great significance to design novel precipitants with high specificity and efficiency for selective conversion of silver ions from wastewater into antibacterial materials.

Schiff bases have proven to be effective probes for metal ions through ligand-metal coordination. (Khan et al., 2021) Although some Schiff bases synthesized by heterocyclic structure containing nitrogen and sulphur showed biological activity after complexing with Ag⁺ ions, (Adeleke et al., 2021; Mary Jenisha Barnabas et al., 2021) they were rarely employed in the direct conversion of Ag⁺ ions from wastewater into antibacterial materials. This is due to the challenge of specific complexation with Ag⁺ ions and formation of precipitated antibacterial material simultaneously.

Herein, we reported a planar benzothiazole Schiff base derivative BTS1 for one-step converting Ag⁺ ions from wastewater into antibacterial material via highly selective chemical precipitation (Fig. 1a). In comparison with twisted molecules, the molecule adopting planar structure is aimed to reduce the solubility, and thus is beneficial to enhance precipitating efficiency. Benzothiazole is introduced for its inherent conjugate rigid plane structure and as building block for antibacterial materials. (Zeng et al., 2019; Yan et al., 2020) Besides, the lone pair of electrons on the nitrogen atom of the thiazole ring would enhance the affinity with Ag⁺ ions. **BTS1** was synthesized via one-pot condensation of 4-(thiophen-2-yl)benzaldehyde (1-CHO) and 2-hydrazinobenzothiazole (HBT) in a high yield up to 82% (Fig. 1b). To demonstrate that the planar molecular structure is superior in precipitation, two control benzothiazole Schiff base derivatives BTS2 and BTS3 with decreased planarity sequentially were synthesized by replacing 1-CHO with [2,2'-bithiophene]-5-carbaldehyde (2-CHO) and 4-(diphenylamino)benzaldehyde (3-CHO), respectively. The precipitation efficiency was decreased from BTS1 to BTS2, and no precipitation was observed in the case of highly twisted BTS3. Remarkably, the precipitate BTS1-Ag exhibited antimicrobial activity against both gram-positive (*S. aureus*) and gram-negative (*E. coli* and *P. aeruginosa*) bacteria. This work shows the highly potential application of Schiff bases in converting precious metals from wastewater into useful materials.

2. Results and discussion

To understand the configurations of the Schiff bases, single crystals of BTS1-3 were cultivated from the binary mixture of THF/methanol (v/v, 2/1) at 25 °C (crystallographic data in detail were provided in Tables S1-S3). As shown in Fig. 2a, BTS1 appeared in an approximately planar configuration from side view with the dihedral angles between the phenvl ring and the adjacent benzothiazolvl and thienvl groups of 3.6° and 3.3°, respectively. BTS2 exhibited a slightly twisted configuration with the dihedral angles increased to 7.8° and 5.0°, respectively (Fig. 2b). For BTS3, the triphenylamine moiety presented highly twisted configuration and the dihedral angle between the phenyl ring and benzothiazolyl unit was up to 22.3°, as shown in Fig. 2c. These results indicated that their solubilities would increase from BTS1 to BTS3 due to the increased spatial configuration of the molecules. Besides, three types of intermolecular interactions including S2...S2 (3.514 Å), N1...N2-H (2.946 Å) and N1...H-N2 (2.142 Å) were detected between BTS1 molecules (Fig. 2d). Assisted by such interactions, planar BTS1 molecules formed interlaced herringbone patterns (Fig. 2g). More types of intermolecular interactions between BTS2 molecules and between BTS3 molecules were found (Fig. 2e and f). These interactions caused BTS2 and BTS3 molecules to pack more loosely in face-toface parallel patterns (Fig. 2h and i). Both in terms of molecular configuration and packing mode, BTS1 seemed preferable to form precipitates.

Forming precipitates with Ag⁺ ions in solution via chemical complexation is the first key issue for **BTS1** to recycle Ag⁺ ions effectively from wastewater. So, we studied the precipitation behaviours between BTS1 (25 μ M) and various metal ions (10 mM) in a mixture of $H_2O/acetone (v/v, 99/1)$, respectively. As shown in Fig. 3, before adding metal ions, BTS1 exhibited two absorption bands centered at 311 nm ($\varepsilon = 2.36 \times 10^4 \text{ M}^{-1}$ cm⁻¹) and 363 nm ($\epsilon = 1.05 \times 10^5 \text{ M}^{-1} \text{ cm}^{-1}$), which were ascribed to π - π * transition of the aromatic rings and n- π * transition related to molecular orbitals of C = N group and thiazole ring, respectively. Remarkably, after adding Ag⁺ ions, the absorption intensity at 363 nm dramatically decreased about 86.4 % (Fig. S1a). In this process, a yellow flocculent precipitate emerged rapidly, and the precipitation process completed within seconds (Video S1). However, the absorption peaks remained almost the same after the addition of other metal ions, implying that no complex reaction occurred, or soluble complexes were formed. The results suggest that $BTS1\ \mbox{can}$ selectively precipitate Ag⁺ ions in approximate aqueous solution. To investigate the influence of the molecular planarity on the precipitation efficiency, BTS2 and BTS3 were also tested in the same condition. As shown in Fig. S1b and c, partial decline (51.5%) and nearly no decline (1.3%) of absorbance were found after adding Ag⁺ ions into the solution of BTS2 and BTS3, respectively. Therefore, from the view of molecular

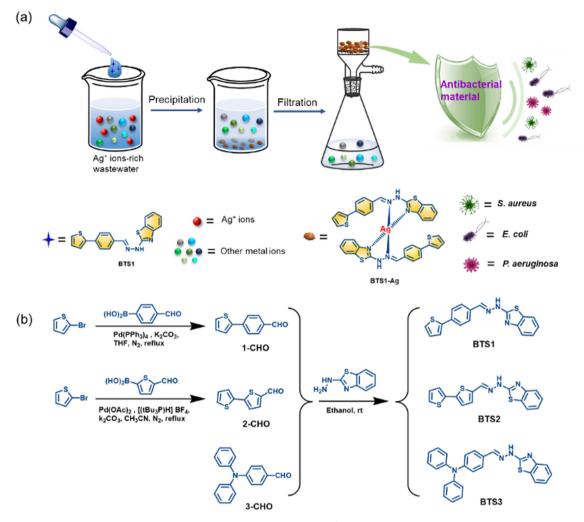


Fig. 1 (a) Graphical illustration of recovery and conversion process of Ag^+ ions from wastewater. (b) The synthetic routes of compound BTS1-3.

configuration, planar structure (**BTS1**) benefitted to precipitate Ag⁺ ions in comparison with nonplanar counterparts (**BTS2** and **BTS3**), which confirmed the feasibility of the proposed molecular design strategy.

The morphology of the yellow precipitates was investigated by scanning electron microscope (SEM) measurement. As shown in Fig. S2a and 2b, BTS1-Ag exhibited aggregated nanosheets morphology, indicating the BTS1-Ag molecules are tending to form aggregates then precipitate from water. The chemical composition and structure of the precipitate were firstly analysed by elemental mapping and X-ray Photoelectron Spectroscopy (XPS). Elemental mapping analyses indicated C, S and Ag elements are uniformly distributed in the precipitate BTS1-Ag (Figure S2c, 2d and 2e). Additionally, the wide-scan XPS spectra indicated that only the yellow precipitate contained Ag3d peak in addition to C1s, S2p, and N1s peaks compared with pure BTS1, which verified the presence of silver element in the precipitate (Fig. 4a). Furthermore, the peaks located at 367.7 and 373.7 eV can be assigned to $Ag3d_{5/2}$ and $Ag3d_{3/2}$ (Fig. 4b), which demonstrated that silver existed in the form of Ag(I). (Zhao et al., 2020) Thus, it was suggested that a chemical complexation reaction existed between **BTS1** and Ag⁺ ion, which yielded a poorly soluble Ag-complex (named as **BTS1-Ag**) and therefore was responsible for the formation of the precipitate and the disappearance of the absorption peaks.

The coordinate mode of BTS1 with Ag⁺ ions in the complex BTS1-Ag was further analysed. As shown in Fig. 5a, the binding energy peaks at 164.0 and 165.3 eV signalled the existence of $S2p_{3/2}$ and $S2p_{1/2}$ orbitals, respectively, which confirmed the confinement of S to C-S bond. (Li et al., 2021) Meanwhile, there were no evident differences in the characteristic peaks of S2p for BTS1 and BTS1-Ag, indicating the S atom has not taken part in complexation of Ag⁺ ions. Notably, the N1s core-level spectrum changed obviously after complexing with Ag⁺ ions. The N1s core-level spectrum of **BTS1** revealed the existence of three components at the binding energy of 398.25, 399.99 and 401.40 eV, attributable to -NHnitrogen atom, thiazole nitrogen atom and the imine nitrogen atom, respectively (Fig. 5b). However, after complexing with Ag⁺ ions, only two components could be detected at the binding energy of 399.22 and 400.03 eV, which can be assigned to the nitrogen atoms of -NH- and N-Ag, respectively (Fig. 5c). (Qian et al., 2020) Therefore, in comparison with BTS1, the disappearance of peaks at 398.25 and 399.99 eV demonstrated the coordination of the imine nitrogen atom

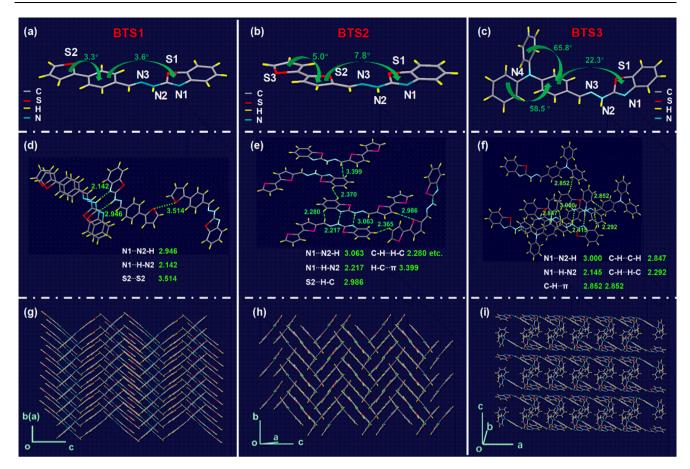


Fig. 2 Molecular structures of (a) BTS1, (b) BTS2, and (c) BTS3. Non-covalent interactions of (d) BTS1, (e) BTS2, and (f) BTS3. Molecular packing of (g) BTS1, (h) BTS2, and (i) BTS3.

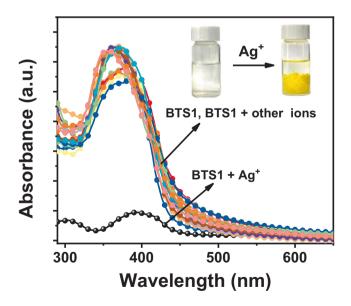


Fig. 3 UV-vis absorption spectra of BTS1 (25 μ M) upon addition of metal ions (10 mM) in H₂O/acetone (v/v, 99/1) (Inset: Changes of BTS1 solution before and after adding Ag⁺ ions under daylight).

and thiazole nitrogen atom with Ag⁺ ions in BTS1-Ag. The FT-IR spectra of BTS1 and BTS1-Ag further verified the binding sites of Ag⁺ ions with BTS1. As shown in Fig. 5d, compared with the FT-IR spectra of BTS1, only the -C = Nabsorption of imine and thiazole was weakened and shifted from 1612 cm^{-1} to 1587 cm^{-1} and 1562 cm^{-1} to 1560 cm^{-1} , respectively. Subsequent experiments were conducted to ascertain the complexation ratio of **BTS1** with Ag⁺ ions. A Job's plot obtained from absorption data showed 2:1 stoichiometric complexation between BTS1 and Ag⁺ ions (Fig. 5e). The 2:1 binding stoichiometry was further confirmed by electrospray ionization (ESI) mass spectra. A peak of [2BTS1 + Ag]⁺ at m/z 778.004 and a peak of $[2BTS1 + Ag + 2H]^+$ at m/z780.008 were found, which agreed with the 2:1 complex of BTS1-Ag (Fig. 5f). Based on the above investigations, the accurate structure of complex BTS1-Ag was proposed as insert picture in Fig. 5e.

To explore the suitable pH for the treatment of Ag^+ ions by **BTS1**, we monitored the changes of the absorption spectra over pH values ranging from 3 to13. As shown in Fig. 6a and Figure S3, for blank **BTS1** (25 μ M), no obvious changes in its absorption profile and absorption intensity at 363 nm were observed in the pH range of 3–11. When pH value was above 12, its absorption intensity was decreased and a new band at 445 nm appeared, probably due to the deprotonation of

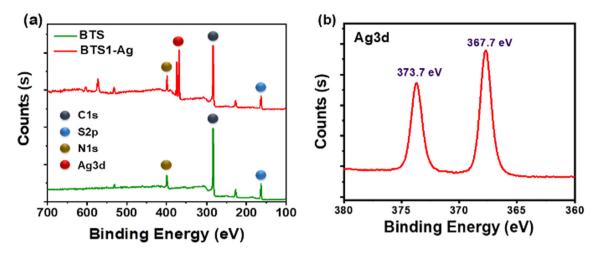


Fig. 4 (a) Wide-scan XPS spectra of BTS1 and BTS1-Ag. (b) High-resolution XPS spectra of Ag3d in BTS1-Ag.

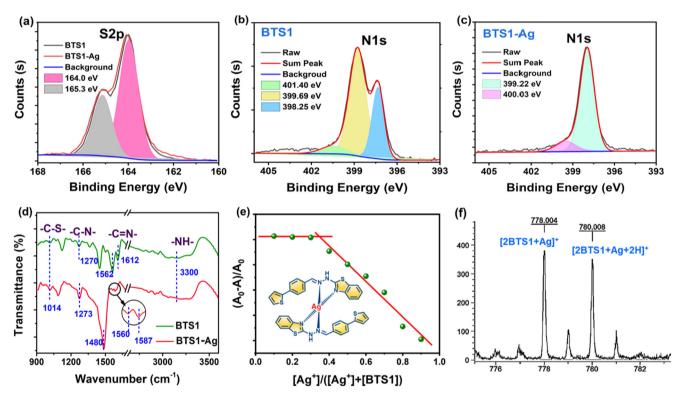


Fig. 5 High-resolution XPS spectra of (a) S2p region in **BTS1** and **BTS1-Ag**, (b) N1s region in **BTS1**, and (c) N1s region in **BTS1-Ag**. (d) FT-IR spectra of **BTS1** and **BTS1-Ag**. (e) Job's plot of **BTS1** towards Ag⁺ ions (Insert: Molecular structure of **BTS1-Ag**). (f) The MS spectrum of **BTS1-Ag**.

BTS1 in the strong alkaline condition. Once a massive overdose of Ag^+ ions (10 mM) was added, the intensity of absorption was dropped obviously accompanied by precipitate formation in the pH range of 6 to 11 (Fig. 6a). The absorption profile remained constant when pH value was below 6, which can be attributed to the competitive protonation effect in acidic condition with the Ag^+ complexation. These results suggested that the optimum pH range for Ag^+ ion precipitation was 6 to 11. Consequently, the following tests were operated by using HEPES buffer (10 mM) at a neutral pH of 7.4. Additionally, the precipitation efficiency exhibited a positive correlation with both **BTS1** concentration and temperature. Precipitation rate (Q) calculated by Equation S1 as a function of temperature was plotted in Fig. 6b. When the initial concentration of **BTS1** was raised to 50 μ M, the Q value reached as high as 95.4% at 50 °C.

The sensitivity of **BTS1** towards Ag^+ ions was further studied through a series of titration experiments at the pH value of 7.4 and the temperature of 25 °C. As shown in Fig. 6c, the absorption intensity at 363 nm was decreased with the increasing amount of Ag^+ ions until 0.5 equiv. of Ag^+ ions were added, which supported the 2:1 complex stoichiometry. The

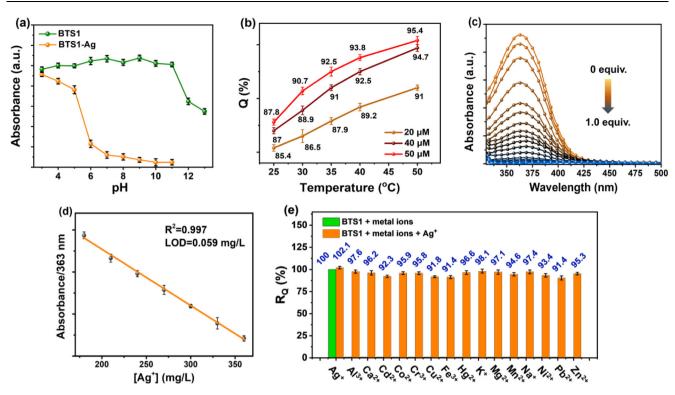


Fig. 6 (a) The absorbance changes of absorption peak at 363 nm for **BTS1** (25 μ M) and **BTS1-Ag** (adding 10 mM of Ag⁺ ions) under different pH conditions. (b) Q variations at different temperatures (25–50 °C) with predetermined **BTS1** concentrations (20 μ M, 40 μ M, 50 μ M). (c) UV–vis absorption spectra of **BTS1** (25 μ M) with various Ag⁺ concentrations (0–1.0 equiv.). (d) The linear relationship at the concentrations of 180–360 mg/L. (e) R_Q variations in the simulated wastewater with different competing ions. Except (a), the rest were all conducted in acetone/H₂O (v/v, 1/99) HEPES buffer (10 mM, pH 7.4).

plot in Fig. 6d displayed a linear relationship between the absorption intensity at 363 nm and the Ag⁺ ion concentration ranging from 180 to 360 mg/L with a linear correlation $R^2 = 0.997$. The corresponding limit of detection (LOD) of **BTS1** toward Ag^+ ions was estimated to be 0.059 mg/L, (Haldar and Lee, 2019) which was much lower than the specified value (0.5 mg/L) in integrated wastewater discharge standard (GB8978-1996). This result indicates that BTS1 is highly sensitive towards Ag⁺ ions. Furthermore, it is crucial for the precipitant to selectively precipitate the specific target from wastewater in the presence of potential competing species. As shown in Fig. 6e, BTS1 exhibited high precipitationefficiency towards Ag⁺ ions in the simulated Ag⁺-rich wastewater, in which the concentration of the competing metal ions was identical to that of Ag^+ ions (100 mg/L). The mass ratio (R_O) between the silver-precipitation with and without competing ions was always within the range of 91.4%-98.1% owing to the high affinity of **BTS1** to Ag^+ ions, despite the existence of the potent competitors Pb^{2+} , Fe^{3+} and Cd^{2+} . The mass loss of precipitation in the presence of other metal ions can be attributed to the formation of soluble complexes between them and BTS1. Considering the obvious merits of fast chemical precipitation and high selectivity, BTS1 should be an outstanding candidate for Ag⁺ ions recovery from wastewater.

Encouraged by the superior antimicrobial property of silver complexes, we evaluated the antimicrobial activity of **BTS1-Ag** against *S. aureus* (gram-positive), and *E. coli*, *P. aeruginosa*

(gram-negative) by using disc diffusion methods. The perceptivity of the bacterial strains can be confirmed from the diameter of the inhibition zone (DIZ). As depicted in Fig. 7a, clear inhibition zones can be observed around the BTS1-Ag-based tested papers, while almost no inhibition zone appeared around BTS1-based and blank test papers, indicating that BTS1-Ag was antimicrobial active for all tested bacterial stains. Under BTS1-Ag stress, E. coli showed a higher antibacterial sensitivity compared with other bacteria. DIZ values were 9.9 \pm 0.72, 13.7 \pm 0.80 and 10.3 ± 0.68 mm for S. aureus, E. coli, and P. aeruginosa, respectively (Fig. 7b and Table S4). We further tested the antibacterial effect of BTS1-Ag by varying its coating concentrations (0- 4.0 g/m^2) on the test papers. As shown in Fig. 7c and 7d, the DIZ values of all bacteria strains were increased dramatically with increasing the concentration of BTS1-Ag in the range of 0 to 1.0 g/m^2 . It was worth noting that **BTS1-Ag** still showed a certain antibacterial effect even at a low concentration of 0.25 g/m^2 with DIZ values of 8.7 ± 0.42 , 9.1 ± 0.38 , 9.5 \pm 0.47 mm for S. aureus, E. coli, and P. aeruginosa, respectively. The antibacterial activity of BTS1-Ag maybe due to its high lipophilicity, which is beneficial to enhance the permeation of the complexes into lipid membranes and the metal binding sites prevent the enzyme activity of the microorganism. (Gönal et al., 2012) Besides, as a by-product in the recovery process, complex BTS1-Ag exhibited superior cost effectiveness in antibacterial field. In this way, BTS1-Ag with admirable antimicrobial activity supported the reutilizing strategy for silver ions from wastes into antibacterial materials.

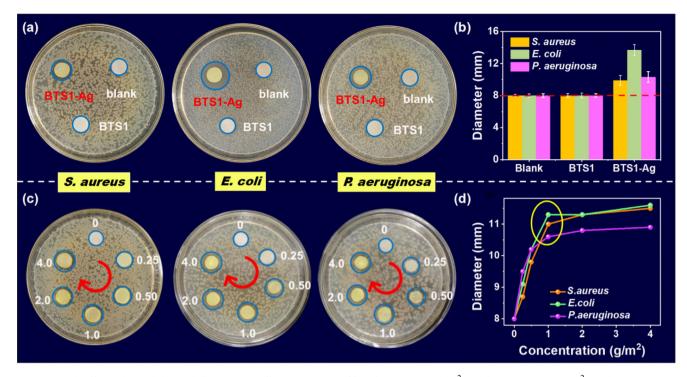


Fig. 7 (a) Antibacterial activity and (b) corresponding DIZ (mm) of blank, **BTS1** (1.0 g/m^2), and **BTS1-Ag** (1.0 g/m^2) against *S. aureus*, *E. coli* and *P. aeruginosa* in culture at 37 °C. (c) Antibacterial activity and (d) corresponding DIZ (mm) of **BTS1-Ag** with different concentrations against *S. aureus*, *E. coli* and *P. aeruginosa*.

3. Conclusions

In summary, we have shown the design and preparation of a benzothiazole Schiff base derivate BTS1 for one-step converting Ag⁺ ions from wastewater into antibacterial material through a selective chemical precipitation process. The lone pair electrons on the nitrogen atoms of imine and thiazole group provided suitable binding sites for Ag⁺ ion coordination, and the planar configuration of **BTS1** further facilitated the precipitation of the complex. These combined advantages enabled BTS1 exhibited outstanding selective precipitation with Ag⁺ ions from wastewater, realizing a high precipitation rate of 95.4% towards Ag⁺ ions at 50 °C. Even in the simulated wastewater containing potential competing metal ions, BTS1 still exhibited super specific precipitation efficiency for Ag⁺ ions over 98%. Most importantly, the precipitate BTS1-Ag can be used as antibacterial material directly, which exhibited good antibacterial activity towards S. aureus, E. coli and P. aeruginosa. In consideration of BTS1 with high selectivity and precipitation rate towards Ag+ ions, as well as the good antibacterial activity of BTS1-Ag, the precipitant BTS1 developed in this study is a promising precipitant for practical recovery and reutilization of Ag⁺ ions from wastewater by one-step.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary material

Supplementary data to this article can be found online at https://doi.org/10.1016/j.arabjc.2023.104836.

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